LETTER

Underestimating the damage: interpreting cetacean carcass recoveries in the context of the *Deepwater Horizon*/BP incident

Rob Williams¹, Shane Gero², Lars Bejder³, John Calambokidis⁴, Scott D. Kraus⁵, David Lusseau⁶, Andrew J. Read⁷, & Jooke Robbins⁸

¹Marine Mammal Research Unit, University of British Columbia, Vancouver, Canada

²Department of Biology, Dalhousie University, Halifax, Canada

³Centre for Fish and Fisheries Research, Cetacean Research Unit, Murdoch University, Western Australia

⁴Cascadia Research Collective, Olympia, WA, USA

⁵New England Aquarium, Boston, MA, USA

⁶School of Biology, Aberdeen University, Aberdeen, Scotland, UK

⁷Nicholas School of the Environment, Duke University, Beaufort, NC, USA

⁸Humpback Whale Studies Program, Provincetown Center for Coastal Studies, Provincetown, MA, USA

Keywords

Anthropogenic impacts; dolphin; *Deepwater Horizon*; Gulf of Mexico; mortality; oil; strandings.

Correspondence

Rob Williams, Current address: Sea Mammal Research Unit, Scottish Oceans Institute, St Andrews Fife KY16 8LB. Tel: +44 (0)1334 462630; Fax: +44 (0)1334 463443. E-mail: rmcw@st-andrews.ac.uk

Received

23 September 2010 Accepted 15 February 2011

Editor Leah Gerber

doi: 10.1111/j.1755-263X.2011.00168.x

Abstract

Evaluating impacts of human activities on marine ecosystems is difficult when effects occur out of plain sight. Oil spill severity is often measured by the number of marine birds and mammals killed, but only a small fraction of carcasses are recovered. The Deepwater Horizon/BP oil spill in the Gulf of Mexico was the largest in the U.S. history, but some reports implied modest environmental impacts, in part because of a relatively low number (101) of observed marine mammal mortalities. We estimate historical carcass-detection rates for 14 cetacean species in the northern Gulf of Mexico that have estimates of abundance, survival rates, and stranding records. This preliminary analysis suggests that carcasses are recovered, on an average, from only 2% (range: 0–6.2%) of cetacean deaths. Thus, the true death toll could be 50 times the number of carcasses recovered, given no additional information. We discuss caveats to this estimate, but present it as a counterpoint to illustrate the magnitude of misrepresentation implicit in presenting observed carcass counts without similar qualification. We urge methodological development to develop appropriate multipliers. Analytical methods are required to account explicitly for low probability of carcass recovery from cryptic mortality events (e.g., oil spills, ship strikes, bycatch in unmonitored fisheries and acoustic trauma).

Introduction

The *Deepwater Horizon*/BP oil spill in the Gulf of Mexico was not only the largest in the US history (Machlis & McNutt 2010) but was also the first to release oil at the sea floor (over 1.5 km below sea level) and to involve the widespread use of dispersants below the surface (Mascarelli 2010). However, many media reports have suggested that the spill caused only modest environmental impacts (Grunwald 2010; Walsh 2010), in part because of a low number of observed wildlife mortalities, especially marine mammals (Unified Area Command from the U.S. Fish and Wildlife Service and the

National Oceanic and Atmospheric Association 2010). Not surprisingly, perhaps, this spill has been compared to other acute environmental disasters, such as the 1989 *Exxon Valdez* oil spill (EVOS). In the case of EVOS, the mortality of sea otters became emblematic of environmental impact, as well as a contentious effort to agree on compensation (Ehrenfeld 1990; Estes 1991). In contrast, the *Deepwater Horizon/BP* event has not left such an iconic symbol of devastation. As of November 7, 2010, "only" 101 cetacean (whale, dolphin, and porpoise) carcasses¹ had been detected across the Northern Gulf

¹ http://www.nmfs.noaa.gov/pr/health/oilspill.htm

of Mexico. The critical issue is, therefore, how to interpret this relatively low number of carcass recoveries in terms of impact to populations. The Gulf of Mexico is a semienclosed subtropical sea that forms essentially one ecosystem with many demographically independent cetacean populations (Mullin & Fulling 2004). Some of these cetacean populations, such as killer whales (*Orcinus orca*), false killer whales (*Pseudorca crassidens*), melonheaded whales (*Peponocephala electra*), and several beaked whale species, appear to be quite small, are poorly studied, or are found in the pelagic realm where they could have been exposed to oil and yet never strand. Small, genetically isolated populations of bottlenose dolphins (*Tursiops truncatus*)could have experienced substantial losses either inshore or offshore.

In an ideal world, one would simply compare postspill to prespill abundance estimates. But, it is rare to have good knowledge of long-term trends in wildlife abundance (Bonebrake et al. 2010). Abundance of many marine mammal populations has been monitored for decades, but the low precision of most cetacean abundance estimates would prevent us from detecting all but the most catastrophic declines using conventional nullhypothesis testing (Taylor et al. 2007b). As a result, it would not be very informative to compare pre- and postspill abundance estimates for populations of cetaceans in the northern Gulf of Mexico. An alternative approach is to count the number of carcasses recovered, acknowledging that these recoveries were subject to a number of processes (e.g., sinking, decaying, scavenging, drifting) that reduce detection probability, and then adjusting the counts upward to estimate total mortality. This is the approach that is commonly taken to estimate the effects of power lines on bird mortality, for example, in which it has been shown in one instance that carcass counts underestimate total mortality by 32% (Ponce et al. 2010). This also appears to be the approach being taken to assess impacts of the Deepwater Horizon Incident on whales and dolphins, with the important caveat that the carcass counts appear to be presented at face value, with no attempt to extrapolate to total mortality (Unified Area Command from the U.S. Fish and Wildlife Service and the National Oceanic and Atmospheric Association 2010; Grunwald 2010; Walsh 2010).

Cetacean carcasses do not necessarily strand along coastlines or remain afloat long enough to be detected at sea. The probability of detecting the death of a marine mammal depends on a wide range of physical and biological factors, including: behavioral responses prior to death, proximity of the carcass to shore (or at-sea observers), decomposition rates and processes, water temperature, wind regime, and local currents (Epperly *et al.* 1996). Cetaceans subject to natural predation would obviously leave no carcass at all. Shore recoveries may be very site-specific, such that the likelihood of a carcass drifting to shore varies with the geography of the coastline itself (Faerber & Baird 2010). As such, "oiled" carcasses detected subsequent to the *Deepwater Horizon*/BP event are expected to represent a small fraction of total mortality in the Northern Gulf of Mexico.

Given the magnitude of the spill and complexity of the response, quantifying the ecological impacts will take a long time. To contribute to this effort, we examined historical data from the Northern Gulf of Mexico to evaluate whether cetacean carcass counts in this region have previously been reliable indicators of mortality, and may therefore accurately represent deaths caused by the *Deepwater Horizon*/BP event.

Methods

We estimated historical carcass-detection rates for 14 species of cetaceans in the northern Gulf of Mexico for which species-specific estimates of abundance (Waring et al. 2009a, b; Mullin & Fulling 2004) species-level adultsurvival rates (Taylor et al. 2007a), and stranding records exist (Waring et al. 2009a, b). Estimates of mortality were generated for each species by multiplying recent abundance estimates by the species-specific mortality rate. An annual carcass-recovery rate was then estimated by dividing the mean number of observed strandings each year by our estimate of annual mortality (Table 1). First, an overall pooled carcass-recovery rate was calculated across all cetacean species (n = 14) in the Gulf of Mexico for which data was available by using the expected number of deaths across all species and the total number of observed carcasses across all species. Next, species-specific carcassrecovery rates were calculated using only species-specific values and a mean (n = 14) of those was taken across species. Species-specific carcass-recovery estimates were not generated for Bryde's whale (Balaenoptera brydei), bottlenose dolphins (Tursiops sp.), and Fraser's dolphins (Lagenodelphis hosei) due to uncertainties in their abundance and/or population structure (Waring et al. 2009b). No attempt was made to estimate carcass-recovery rates in the two taxonomic groups that are not identified to species in the field during raw data collection: Kogia (a pooled estimate for two species, dwarf and pygmy sperm whales), or mesoplodonts (a pooled estimate for a genus of similar-looking beaked whales) (Waring et al. 2009b).

Results

Our analysis suggests that an average of 4,474 individual cetaceans died annually between 2003 and 2007 from all

Table 1	Population paramete	rs and illustrative species	s-specific carcass-recove	ery rates for 14 s	species from the Gulf of Mexico

Northern Gulf of Mexico population	Population estimate ^a	CV ^a	$\begin{array}{c} Adult-\\ survival\ rate^b \end{array}$	Estimated annual mortality ^c	Mean observed annual strandings ^a	Carcass-detection rate (%)
Sperm whale	1665	0.20	0.986	23.3	0.8	3.4
Cuvier's beaked whale	65	0.67	0.95	3.3	0.2	6.2
Atlantic spotted dolphin ^d	37611	0.28	0.95	1880.6	2.4	0.13
Pantropical spotted dolphin	34067	0.18	0.95	1703.4	0.8	0.05
Striped dolphin	3325	0.48	0.95	166.3	0.8	0.48
Spinner dolphin	1989	0.48	0.95	99.5	1	1.0
Rough-toothed dolphin ^d	2653	0.42	0.95	132.7	5.8	4.4
Clymene dolphin	6575	0.36	0.95	328.8	0.6	0.18
Killer whale	49	0.77	0.99	0.5	0	0
False killer whale	777	0.56	0.99	7.8	0	0
Pygmy killer whale	323	0.60	0.95	16.2	0.2	1.2
Melon-headed whale	2283	0.76	0.99	22.8	1.4	6.1
Risso's dolphin	1589	0.27	0.95	79.5	2.8	3.5
Short-finned pilot whale	716	0.34	0.986	10.0	0.2	2.0
Average of all species						2.0
Pooled across all species ($n = 14$)	93,687	-	-	4,474	17	0.4

^aPopulation abundance and stranding data (2003–2007) were taken from (Waring et al. 2009b), unless otherwise noted.

^bPopulation-level estimates are preferable but generally unavailable, so data taken from (Taylor *et al.* 2007a).

^cCalculated as the abundance multiplied by mortality rate (= 1-survival rate).

^dPopulation abundance and stranding data (2002–2006) were taken from (Waring et al. 2009a).

natural and anthropogenic causes. However, during that period, an average of only 17 cetacean carcasses were detected annually along the northern Gulf of Mexico. This would suggest that the overall pooled rate of carcass recovery for cetaceans in the Gulf of Mexico is approximately 0.4% of the total estimated mortality. Table 1 breaks down the recovery rates by species. Carcasses were recovered only from a mean of 2.0% (range: 0-6.2%) of cetacean species deaths along the northern Gulf of Mexico. The disparity between this value and the overall pooled value likely results from undue influence of poorly studied and relatively rare species (e.g., Cuvier's beaked whale and melon-headed whale; Table 1) with high estimated carcass-recovery rates that are weighted equivalently and treated as reliably in this average as estimates from species that are common and well studied. We have reason to believe that the Cuvier's beaked whale recovery rate is positively biased. The original abundance estimate is thought to be an underestimate by a factor of 2 to 4, based on the assumption of certain track line detection (Mullin & Fulling 2004). Our carcass-recovery rate for deep-diving whales would then be biased high by a factor of 2–4.

Discussion

Our results indicate that carcass-recovery rates are historically low for cetaceans in the Gulf of Mexico. Studies of other populations show similar recovery rates. In long-term studies of killer whales off the coasts of British Columbia and Washington State, in which populations are censused completely every year, carcasses from confirmed deaths of known individuals are recovered only 6% of the time (Fisheries and Oceans Canada 2008). Similarly, low-detection rates have been estimated for carcasses of eastern gray whales (Eschrichtius robustus, <5%, Heyning & Dahlheim 1990), North Atlantic right whales (Eubalaena glacialis, 17%, Kraus et al. 2005), and harbor porpoises (Phocoena phocoena, <1%, Moore & Read 2008), all of which occur in near-shore waters. Beached carcasses of other pelagic marine vertebrates have been shown to be equally poor indicators of mortality (for example, 7-13% recovery rates for four species of sea turtle, Epperly et al. 1996). As such, raw carcass counts alone are not reliable indicators of the magnitude of mortality for these species.

We do not claim to have calculated definitive multipliers for this spill. Instead, our aim is to show plausible ranges for those multipliers, in order to illustrate how much they would affect our perception of the ecological damages caused by *Deepwater Horizon* incident and why this topic is worthy of additional resources for methodological development. Consider, for example, one sperm whale being detected as a carcass, and a necropsy identified oiling as a contributing factor in the whale's death. If the carcass-detection rate for sperm whales is 3.4% (Table 1), then it is plausible that 29 sperm whale deaths represents the best estimate of total mortality, given no additional information. If, for example, 101 cetacean carcasses were recovered overall, and all deaths were attributed to oiling, the average-recovery rate (2%) would translate to 5,050 carcasses, given the 101 carcasses detected (Table 1). As the necropsy results emerge, we can evaluate whether this prediction is high or low, but the sheer scope for underestimation builds a compelling case, in our view, for additional work. The vast majority of carcasses recovered appear to have been bottlenose dolphins.¹ As necropsy results emerge and the need for recovery plans debated, we encourage such discussions to explicitly take into account the probability that the number of dolphins stranded represented something on the order of only 2% of the number of animals killed. The potential is high for the spill to have caused catastrophic impacts on small, localized populations of bottlenose dolphins in the Gulf. We note that coastal and offshore forms of bottlenose dolphins are found off California, with the coastal carcass having a 50-fold greater probability of stranding than an offshore one (Perrin et al. 2010).

Even in the case of EVOS, the large number of observed deaths was acknowledged to represent only a fraction of the total mortality (Estes 1991). Two approaches were taken to estimate total mortality in Prince William Sound: (1) a comparison of pre- and postspill population size; and (2) extrapolations from recovered carcasses to total mortality from a multiplier based on the probability of recovering a carcass (Garshelis 1997). Our estimates of carcass-recovery rates were calculated from the best available data, but we caution against using historic (i.e., pre-spill) carcass-recovery rates to generate a simple multiplier to assess total mortality in the Deepwater Horizon/BP Incident. On the one hand, considerable efforts were expended by government agencies and others to search for marine mammal carcasses after the spill, which could raise recovery rates above those estimated here. Fortunately, a comparison of pre- and postspill search effort ought to be among the most tractable factors to account for when calculating carcass-recovery rates. On the other hand, there are several arguments to suggest that our carcass-recovery rates are biased high. First, we estimated the number of carcasses using adultsurvival rate; had we included juvenile and calf mortality, the total number of carcasses would have been substantially higher and our estimated carcass-recovery rate substantially lower. The point estimate is strongly influenced by some optimistic values for Cuvier's beaked whale and melon-headed whale (Table 1). Abundance of these elusive species is biased low, due to well-known difficulties in estimating track line detection probability (g(0)) for deep-diving species. Some of these cetaceans represent prey species: our denominators include animals that would have been preyed upon and not ended up as

carcasses. Given that many cetaceans are highly social, entire clusters, schools, pods, matrilines, or groups of animals could have been affected (Williams et al. 2009). Although we used recent population estimates, it has yet to be determined how many animals in each population were actually exposed to the spill. Finally, the location of the spill and the subsequent response effort likely affected the probability of detecting associated deaths. These are the factors that must be carefully considered as efforts to assess population impacts continue. We present our historic recovery rates as starting points for discussion, but caution that incorrect multipliers may result in estimated mortalities exceeding the number of animals that were ever in the vicinity of the spill (Parrish & Boersma 1995). Estimating the correct multipliers will require an interdisciplinary research effort to combine oceanographic and cetacean habitat modeling to assess exposure risk and likely deaths caused by exposure. This research is needed, but currently lacking from research priorities emerging from the oil spill mitigation and recovery efforts.

The issue of carcass-detection rates is not merely of academic interest. Our results are directly relevant to assessment of ecological damages caused by the Deepwater Horizon/BP oil spill, but also have global relevance for litigation and marine conservation policy. Given that environmental restitution in the United States can be based on a violation system (Alexander 2010), carcassrecovery rates must be explicitly considered when evaluating the impacts of such disasters. In the case of EVOS, legal damages placed the value of each sea otter killed at US\$80,000, or the cost of rehabilitating each oiled otter (Estes 1991; Garshelis 1997). In terms of broader recommendations for marine policy, we note that carcass counts are used in many countries, including the United States, to monitor human impacts on cetacean populations. The tools that managers use in the United States to estimate and limit the impacts of human activities on stocks relies upon "potential biological removal" (PBR), a calculation that determines how many animals can be removed from a stock before causing harm. The PBR estimate, under the Marine Mammal Protection Act (MMPA) depends on reasonably unbiased and precise estimates of human-caused mortality (Wade 1998). In contrast, the effects of many human impacts are only witnessed opportunistically, such as a carcass being discovered on a beach. The issue arises when policymakers, legislators, or biologists treat these carcass-recovery counts as though they were complete counts or parameters estimated from some representative sample, when in fact, they are opportunistic observations. Our study suggests that these opportunistic observations should be taken to estimate only the bare minimum number of human-caused mortalities. This work suggests that carcass counts alone

are unreliable indicators of either natural or anthropogenic sources of mortality. It is vital to develop a framework that explicitly accounts for the low probability of recovering carcasses, if we are to accurately assess the sustainability of all cryptic forms of human-caused mortality.

Human impacts on marine ecosystems and marine mammals are growing both in type and scale (Kraus & Rolland 2007; Clausen & York 2008; Duce et al. 2008; Doney 2010; Hoegh-Guldberg & Bruno 2010; Tittensor et al. 2010). Establishing the proper spatial and temporal scales, at which to assess the impacts of acute events, is further complicated by potential long-term effects and a lack of basic population-specific information (Bejder et al. 2006). This highlights the need for long-term population monitoring, such as that mandated by the U.S. MMPA (Bonebrake et al. 2010). In the first year after the 1989 Exxon Valdez spill, the AT1 group of "transient" killer whales experienced a 41% loss; there has been no reproduction since the spill (Matkin et al. 2008). Although the cause of the apparent sterility is unknown, the lesson serves as an important reminder that immediate death is not the only factor that can lead to long-term loss of population viability. The recent disaster in the Gulf of Mexico provides an important opportunity to assess whether or not the intensity of monitoring conducted in the Gulf of Mexico is sufficient to detect even catastrophic effects (Taylor et al. 2007b). If line-transect survey data are found wanting, we see value in exploring new passive acoustic monitoring methods to detect trends in relative abundance (Marques et al. 2009, Rojas-Bracho et al. 2010). These could be especially useful for rare or pelagic species, or those for which g(0) estimation is particularly problematic. Accurate assessment of impacts also must consider how species are likely impacted, whether acutely on contact with oil or over the longer term through toxicity or habitat degradation (Lovett 2010; Schrope 2010b). If support for longer term assessments dwindles as the time passes and public attention moves elsewhere (Schrope 2010a), then chronic effects will remain unknown. In such cases, only immediately observable effects, such as the number of carcasses, have and will be used to determine the impact of an event, and synergistic and lagged effects will not be considered. Our findings suggest that assessments of the impact of anthropogenic events based solely on the numbers of carcasses recovered are deceptively biased. A better understanding of carcass-recovery rates and the degree to which they underestimate actual mortality, is critical to assessing the true consequences of oil spills and other human activities known to cause cryptic mortality, such as ship strikes, certain fisheries interactions, and acoustic trauma.

Acknowledgments

Lynne Barre, Dee Boersma, and Dave Thompson gave valuable comments at an early stage of the development of this manuscript. We thank Leah Gerber, Tim Gerrodette and Barb Taylor for their careful reviews.

References

- Alexander, K. (2010) The 2010 oil spill: criminal liability under wildlife laws. Congressional Research Service Report to Congress, Washington D.C., **R41308**, p. 10.
- Bejder, L., Samuels, A., Whitehead, H., Gales, N. (2006)
 Interpreting short-term behavioural responses to disturbance within a longitudinal perspective. *Anim Behav* 72, 1149–1158.
- Bonebrake, T.C., Christensen, J., Boggs, C.L., Ehrlich, P.R. (2010) Population decline assessment, historical baselines and conservation. *Conserv Lett* **3**, 371–378.
- Clausen, R., York, R. (2008) Economic growth and marine biodiversity: influence of human social structure on decline of marine tropic levels. *Conserv Biol* **22**, 458–466.
- Doney, S.C. (2010) The growing human footprint on coastal and open-ocean biochemistry. *Science* **328**, 1512–1516.
- Duce, R.A., LaRoche, J., Altieri, K., *et al.* (2008) Impacts of atmospheric anthropogenic nitrogen on the open ocean. *Science* **320**, 893–897.
- Ehrenfeld, D. (1990) The lessons of Valdez. Conserv Biol 4, 1–2.
- Epperly, S.P., Braun, J., Chester, A.J., *et al.* (1996) Beach strandings as indicators of at-sea mortality of sea turtles. *Bull Mar Sci* **59**, 289–297.
- Estes, J.A. (1991) Catastrophes and conservation: lessons from sea otters and the *Exxon Valdez*. *Science* **254**, 1596.
- Faerber, M.M., Baird, R.W. (2010) Does the lack of observed beaked whale strandings in military exercise areas mean no impacts have occured? A comparison of stranding and detection probabilities in the Canary and main Hawaiian islands. *Mar Mamm Sci* **26**, 602–613.
- Fisheries and Oceans Canada. (2008) *Recovery strategy for the northern and southern resident killer whales (Orcinus orca) in Canada*. Fisheries and Oceans Canada, Ottawa, Canada.
- Garshelis, D.L. (1997) Sea otter mortality estimated from carcasses collected after the Exxon Valdez oil spill. *Conserv Biol* **11**, 905–916.
- Grunwald, M. (2010) The BP spill: has the damage been exaggerated? *Time Magazine*. Available from http://www. time.com/time/nation/article/0,8599,2007202,00.html. Accessed 4 October 2010.
- Heyning, J.E., Dahlheim, M.E. (1990) Strandings and incidental takes of gray whales. Report to the International. Whaling Commission, SC/A90/G2, 16 pp.
- Hoegh-Guldberg, O., Bruno, J.F. (2010) The impact of climate change on the world's marine ecosystems. *Science* **328**, 1523–1528.

Kraus, S.D., Brown, M.W., Caswell, H., *et al.* (2005) North Atlantic right whale in crisis. *Science* **309**, 561–562.

Kraus, S.D., Rolland, R.M. (2007) The urban whale syndrome. Pages 488–513 in S.D. Kraus, R.M. Rolland, editors. *The urban whale: North Atlantic right whales at the crossroads*. Harvard University Press, Cambridge, Massachusetts.

Lovett, R.A. (2010) Oil spills toxic trade-off. Nature News Online. Available from: http://www.nature.com/ news/2010/101110/full/news.2010.597.html. Accessed 11 November 2010.

Machlis, G.E., McNutt, M.K. (2010) Scenario-building for the deepwater horizon oil spill. *Science* **329**, 1018–1019.

Marques, T.A., Thomas, L., Ward, J., DiMarzio, N., Tyack, P.L. (2009) Estimating cetacean population density using fixed passive acoustic sensors: an example with Blainville's beaked whales. *J Acoust Soc Am* **125**, 1982–1994.

Mascarelli, A. (2010) Deepwater horizon: after the oil. *Nature* **467**, 22–24.

Matkin, C.O., Saulitis, E.L., Ellis, G.M., Olesiuk, P., Rice, S.D. (2008) Ongoing population-level impacts on killer whales Orcinus orca following the 'Exxon Valdez' oil spill in Prince William Sound, Alaska. *Mar Ecol Prog Ser* **356**, 269–281.

Moore, J.E., Read, A.J. (2008) A bayesian uncertainty analysis of cetacean demography and bycatch mortality using age-at-death data. *Ecol Appl* **18**, 1914–1931.

Mullin, K.D., Fulling, G.L. (2004) Abundance of cetaceans in the oceanic northern Gulf of Mexico, 1996–2001. *Mar Mamm Sci* 20, 787–807.

Parrish, J.K., Boersma, P.D. (1995) Muddy waters. *Amer Sci* **83**, 112–115.

Perrin, W.F., Thieleking, J.L., Walker, W.A., Archer, F.I., Robertson, K.M. (2010) Common bottlenose dolphins (Tursiops truncatus) in California waters: Cranial differentiation of coastal and offshore ecotypes. *Mar Mamm Sci*, doi: 10.1111/j.1748-7692.2010.00442.x.

Ponce, C., Alonso, J.C., Argandona, G., Garcia Fernandez, A., Carrasco, M. (2010) Carcass removal by scavengers and search accuracy affect bird mortality estimates at power lines. *Anim Conserv* **13**, 603–612.

Rojas-Bracho, L., Jaramillo-Legoretta, A., Cardenas, G., et al. (2010) Assessing trends in abundance for vaquita using acoustic monitoring: within refuge plan and outside refuge research needs. Workshop Report—October 19–23, 2009. U.S. Department of Commerce, NOAA Technical Memorandum NMFS, NOAA-TM-NMFS-SWFSC-459. 39 pages.

Schrope, M. (2010a) The lost legacy of the last great oil spill. *Nature* **466**, 304–305.

Schrope, M. (2010b) Oil spill cruise finds a field of dead coral. Nature News Online. Available from: http://www.nature. com/news/2010/101105/full/news.2010.589.html Accessed November 11 2010.

Taylor, B.L., Chivers, S.J., Larese, J., Perrin, W.F. (2007) Generation length and percent mature estimates for IUCN assessments of cetaceans. Administrative Report LJ-07-01, Southwest Fisheries Science Center, 8604 La Jolla Shores Blvd., La Jolla, CA 92038, USA. 24 pp.

Taylor, B.L., Martinez, M., Gerrodette, T., Barlow, J., Hrovat, Y.N. (2007b) Lessons from monitoring trends in abundance of marine mammals. *Mar Mamm Sci* 23, 157–175.

Tittensor, D.P., Mora, C., Jetz, W., *et al.* (2010) Global patterns and predictors of marine biodiversity across taxa. *Nature* **466**, 1098–1101.

Unified Area Command from the U.S. Fish and Wildlife Service and the National Oceanic and Atmospheric Association. (2010) *Deepwater/Horizon* response consolidated fish and wildlife collection report. Available from: http://www.restorethegulf.gov/sites/default/files/ Consolidated%20Wildlife%20Table%20100410.pdf. Accessed 5 October 2010.

Wade, P.R. (1998) Calculating limits to the allowable human-caused mortality of cetaceans and pinnipeds. *Mar Mamm Sci* 14, 1–37.

Walsh, B. (2010) Oil spill: the well is dead. *Time Magazine*.Available from: http://ecocentric.blogs.time.com/2010/ 09/19/oil-spill-the-well-is-dead/. Accessed 4 October 2010.

Waring, G.T., Josephson, E., Fairfield-Walsh, C.P., Maze-Foley, K. (2009a) U.S. Atlantic and Gulf of Mexico marine mammal stock assessments – 2008. NOAA Tech Memo NMFS NE 210, p. 440.

Waring, G.T., Josephson, E., Maze-Foley, K., Rosel, P.E. (2009b) U.S. Atlantic and Gulf of Mexico marine mammal stock assessments – 2009. NOAA Tech Memo NMFS NE 213, p. 528.

Williams, R., Lusseau, D., Hammond, P. (2009) The role of social aggregations and protected areas in killer whale conservation: the mixed blessing of critical habitat. *Biol Conserv* 142, 709–719.